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**Separation and
attribution of aerosol
absorption**

M. Yang et al.

Attribution of aerosol light absorption to black carbon, brown carbon, and dust in China – interpretations of atmospheric measurements during EAST-AIRE

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Abstract

Black carbon, brown carbon, and mineral dust are three of the most important light absorbing aerosols. Their optical properties differ greatly and are distinctive functions of the wavelength of light. Most optical instruments that quantify light absorption, however, are unable to distinguish one type of absorbing aerosol from another. It is thus instructive to separate total absorption from these different light absorbers to gain a better understanding of the optical characteristics of each aerosol type. During the EAST-AIRE (East Asian Study of Tropospheric Aerosols: an International Regional Experiment) campaign near Beijing, we measured light scattering using a nephelometer, and light absorption using an aethalometer and a particulate soot absorption photometer. We also measured the total mass concentrations of carbonaceous (elemental and organic carbon) and inorganic particulates, as well as aerosol number and mass distributions. We were able to identify periods during the campaign that were dominated by dust, biomass burning, fresh (industrial) chimney plumes, other coal burning pollution, and relatively clean (background) air for Northern China. Each of these air masses possessed distinct intensive optical properties, including the single scatter albedo and Ångström exponents. Based on the wavelength-dependence and particle size distribution, we apportioned total light absorption to black carbon, brown carbon, and dust; their mass absorption efficiencies at 550 nm were estimated to be 9.5, 0.5, and 0.03 m²/g, respectively. While agreeing with the common consensus that BC is the most important light absorber in the mid-visible, we demonstrated that brown carbon and dust could also cause significant absorption, especially at shorter wavelengths.

1 Introduction

Aerosols scatter and absorb shortwave solar radiation, generally resulting in cooling and warming of the planet, respectively, if the two phenomena are treated independently. When globally averaged, the vast amounts of highly scattering sulfate particles

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in the atmosphere (a product of fossil-fuel combustion) are believed to have a net cooling effect on the earth's surface. On the other hand, strongly absorbing soot carbon aerosols lead to a warming of the planet (Solomon et al., 2007). Other important light absorbing aerosols in the atmosphere include brown carbon and mineral dust.

5 Aerosol absorption can be represented by the classic Beer-Lambert Law, which relates the intensity of incident (I_0) and outgoing (I) light by an exponential:

$$I = I_0 e^{-\alpha l c} \quad (1)$$

where α is the absorption cross section, l is the path length, and c is the concentration. The product of α and c in the exponent is defined as the absorption coefficient (b_{ap}), which can be measured by in-situ instruments such as the aethalometer in units of inverse length. If c represents the mass concentration of absorbing particles (e.g. $\mu\text{g}/\text{m}^3$), α is then the mass absorption cross-section, or mass absorption efficiency (MAE), often in units of m^2/g ; MAE can vary greatly depending on the type of light absorbers. Analogously, the scattering coefficient (b_{sp}) is the product of the scattering cross section and the concentration of scattering particles. Since all aerosols scatter light, scattering coefficient is often used as a proxy for particle concentration.

15 The fraction of light extinction (sum of scattering and absorption) due to scattering is defined as the single scatter albedo (ω_0 , or SSA) – an intensive property determined by the particle composition while independent of the total aerosol concentration:

$$20 \quad \omega_0 = b_{sp} / (b_{ap} + b_{sp}) \quad (2)$$

The single scatter albedo is an important factor in controlling whether an aerosol layer has a cooling or warming effect. Non-absorbing particles (e.g. ammonium sulfate) have ω_0 of unity, whereas a lower ω_0 implies more absorption.

25 Aerosols scatter and absorb light differently depending on the particle size (D_p) relative to the wavelength of light (λ). For a given aerosol composition (which determines the refractive index), the Mie scattering theory can be used to approximate the amount of scattering and absorption when D_p and λ are known, assuming the particle is spherical. Since most atmospheric aerosols are small, their absorption tends to be higher

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near the ultraviolet (UV) than near the infrared (IR). Absorption is often parameterized as proportional to $\lambda^{-\tilde{A}}$, where the exponent, \tilde{A} , is the absorption Ångstrom exponent. An analogous expression can be written for the wavelength-dependence of scattering.

Soot carbon, or combustion-generated carbonaceous aerosol that absorbs light strongly over the entire visible spectrum, is traditionally viewed as the most important light absorber for solar radiation (Hansen et al., 1984; Penner et al. 1993). Soot is monitored either optically or chemically. The optical method relies on light absorption; the corresponding operational definition for soot is black carbon (BC). The chemical (thermal) method measures the mass concentration of total carbon, in which the portion that is refractory and does not volatilize below $\sim 400^\circ\text{C}$ in air is termed elemental carbon (EC) (Birch and Cary, 1996); the volatile portion is assumed to be organic carbon (OC). Since BC and EC refer to somewhat different fractions of absorbing carbon, distinctions in these operational definitions need to be carefully considered when relating optical and chemical measurements of soot, such as when calculating the mass absorption efficiency.

Freshly emitted soot particles from combustion have graphite-like sheet structures that fold among themselves to form agglomerates of primary spherules around 20 nm in diameter (Bond and Bergstrom, 2006), which shortly after emission coagulate and form loose aggregates. Because they are usually small, soot particles fit within the Rayleigh scattering regime for near-visible wavelengths. In this regime, absorption follows a theoretical λ^{-1} relationship, implying a wavelength-independent imaginary component of the refractive index; this for soot is supported by the Band-gap theory (Bond and Bergstrom, 2006). Many laboratory observations have shown that soot has an absorption Ångstrom exponent of one in the solar spectrum (Bergstrom et al., 2002; Schnaiter et al., 2003; Kirchstetter et al., 2004). Previous experiments from our group also demonstrated that within the range measured by the seven-wavelength aethalometer (370 to 950 nm), $\tilde{A} \approx 1$ was obtained when fitting absorption with a power law for air masses dominated by soot carbon (Kline et al., 2004).

Modeling studies have suggested that that total soot absorption is increased by

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~30% as a result of aggregation from singular primary spherules (Fuller et al., 1999), possibly due to the effect of multiple scattering. Bond and Bergstrom (2006) amassed literature results and recommended a mass absorption efficiency of $7.5 \pm 1.2 \text{ m}^2/\text{g}$ for fresh soot aggregates at 550 nm. Other aerosols can also be externally or internally mixed with soot. Fuller et al. (1999) and Bond et al. (2006) have shown that a complete encapsulation of a soot core with non-absorbing organic or inorganic condensates might result in an absorption enhancement of up to 30~50%.

More recently, carbonaceous particles that are optically in between the strongly absorbing soot carbon and non-absorbing organic carbon have also been shown to be significant, and they are operationally defined as brown carbon. Formed by inefficient combustion of hydrocarbons (e.g. smoldering) and also by photo-oxidation of biogenic particles, brown carbon encompasses a large and variable group of organic compounds, and may include humic substances, polyaromatic hydrocarbons, and lignin (Andreae and Gelencser, 2006). Due to the presence of resonant ring structures, most absorption by brown carbon takes place in the UV, though the lowered bond energy of the conjugated bonds causes a tail of absorption in the short visible wavelengths. Consequently, the absorption Ångstrom exponent of brown carbon is larger than that of soot.

Thus far, published absorption efficiencies and Ångstrom exponents for “brown carbon” have been inconsistent. Hoffer et al. (2006) extracted humic-like substances (HULIS) with water from the fine fraction of biomass burning aerosols and found an absorption Ångstrom exponent of 6~7 and a mass absorption efficiency of $\sim 0.03 \text{ m}^2/\text{g}$ at 532 nm. Kirchstetter and Novakov (2004) isolated organic carbon with acetone and found mass absorption efficiencies of $0.9 \text{ m}^2/\text{g}$ at 500 nm and $0.6 \text{ m}^2/\text{g}$ at 550 nm for the extract ($\text{Å} > 4$ over this short range of wavelengths). Clarke et al. (2007) isolated “refractory organic carbon” from biomass burning aerosols with a thermal method and approximated brown carbon absorption as the difference between total and extrapolated BC absorption; from these estimates they obtained mass absorption efficiencies of $0.52 \text{ m}^2/\text{g}$ at 470 nm and $0.19 \text{ m}^2/\text{g}$ at 530 nm ($\text{Å} > 8$ over this short range of wave-

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lengths). It is unsurprising to see such discrepancies in literature results because a variety of aerosols were measured using several different techniques, with each derived concentration and light absorption corresponding to a different fraction of carbonaceous particles.

5 Optical properties of eolian dust aerosols are determined by their mineralogical origins as well as physical modifications once airborne. Ferric iron oxides such as hematite and goethite, when internally mixed with clay minerals, result in significant dust absorption in the UV/visible (Sokolik and Toon, 1999). As with brown carbon, dust absorption in the solar spectrum decreases rapidly with increasing wavelength.
10 Alfaro et al. (2004) found that for Gobi (China), Sahara (Tunisia), and Sahel (Niger) dust, the mass absorption efficiencies ranged from $0.01\sim0.02\text{ m}^2/\text{g}$ at 660 nm and $0.06\sim0.12\text{ m}^2/\text{g}$ at 325 nm ($\text{\AA}\approx 3$ over this range), while an absorption Ångström exponent of 2.6 can be calculated from the report of Fialho et al. (2006).

Particles generated from diesel combustion are conventionally used as the laboratory surrogate for black carbon. Much less is known about coal-burning aerosols, even
15 though coal combustion accounts for roughly half of the BC in the atmosphere (Bond et al., 2002). The EAST-AIRE campaign in 2005 provided us an opportunity to study the optical and chemical properties of, among others, coal-derived aerosols, which are ubiquitous in much of China and other developing countries. Since coal generally contains more impurities and water than diesel fuel does, we expect the combustion of coal, particularly in residential usage, to be less efficient and more incomplete than the
20 combustion of higher-grade fossil fuels. Consequently, a significant amount of brown carbon might be generated, which is less absorbing than BC per unit mass but has a greater absorption Ångström exponent. A high wavelength-dependence in absorption might also indicate the presence of dust. The broad spectrum of our aethalometer (370
25 to 950 nm) enables us to closely examine these wavelength dependences.

Ambient aerosol absorption is often normalized by the concentration of EC to approximate the EC mass absorption efficiency. While this approach might be valid in areas dominated by fresh soot emission, dividing total absorption by the soot concentration

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is misleading when other light-absorbers are also present. In those more complex environments, if we can separate and attribute total light absorption to different absorbing aerosols, given their respective mass concentrations, we can calculate MAE specific to each aerosol type present.

2 Experimental

EAST-AIRE took place in part between 2 March and 26 March 2005 at the Institute of Atmospheric Physics (IAP) research facility ~2 km east of Xianghe, China (Li et al., 2007). About 70 km east of Beijing, our site was heavily polluted due to both local and distant sources of fossil fuel combustion, vehicular and industrial emissions, and also occasional dust and biomass fires. While coal burning was technically forbidden in the town of Xianghe, nearby residences and a number of small factories neighboring the IAP facility used coal. The boiler of the IAP facility itself burned coal also; a cyclone was installed to remove large particles – an implementation that most other boilers lacked. We could see over 20 active chimneys in all directions from our sampling site. Direct plumes from these chimneys almost always contained particle number concentrations over 10^4 cm^{-3} . Given the density and proximity of sources, back-trajectories to identify specific emissions were not attempted.

The ambient temperature was around 0°C at the start of the campaign and warmed up to $\sim 20^\circ\text{C}$ towards the end. The air was very dry, with a campaign average relative humidity below 40%. Due to its inland geographical location, there was no observable marine influence. Brisk winds ($\sim 10 \text{ m/s}$) usually came from the west/northwest (the direction of Beijing and distant mountain ranges), which brought along relatively clean air and occasionally mineral dust particles. In contrast, much gentler winds from the southeast (the direction of Xianghe town) carried the highest levels of air pollutants. The air was often particularly dirty in the morning when the winds were weak and the inversion layer was low, as emissions from anthropogenic activities including morning traffic, cooking, and heating were trapped near the ground. Sometimes a second (and

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somewhat smaller) peak in atmospheric pollutants was also observed in the evening traffic and cooking hours.

5 A number of different aerosol types are generated from coal combustion, including BC, fly ash (condensed silicates), and unburned coal ash. Brown carbon can also be produced as a result of incomplete combustion (Bond et al., 2002) – a likelihood that was supported by our observation of yellowish smoke from some nearby chimneys. The production of brown carbon might have been enhanced since in many households, furnaces are partially choked on purpose to moderate temperature and prolong burning time (personal experience). While large heating systems usually employ chunks of coal
10 that are unmodified, for residential cooking and heating, honeycomb coal briquettes are preferred because they burn more predictably and produce less smoke. Coal briquettes are made by first binding powdered coal with clay when wet; the mixture is then molded into cylinders with many cylindrical cavities, resembling a honeycomb. After the coal is burned off, briquettes retain their structural integrity and appear orange in color,
15 suggesting that the clay binder contains ferric oxides. Filter from our OCEC carbon aerosol analyzer that were exposed to coal-dominated pollution for ~12 h appeared to have the same orange color after thermal analysis as burned coal briquettes. It is possible that such iron-rich clay materials can be liberated from the briquette matrix as coarse, light absorbing aerosols. The emission of light absorbing particles other than
20 BC might be the cause for the wide range of absorption Ångstrom exponents observed for coal-derived aerosols (1–2.9, Bond et al., 2002).

2.1 Absorption and scattering measurements

We used a Radiance Research (Seattle, WA) Particle Soot Absorption Photometer (PSAP) that recorded absorption at a single wavelength (567 nm) every 5 s and a
25 TSI (St. Paul, MN) Model 3563 integrating nephelometer that recorded scattering at 450, 550, and 700 nm every second. The PSAP and the nephelometer shared an impactor that switched from submicron particle sampling to total particle sampling every 10 min. We adjusted raw scattering coefficients from the nephelometer for angular

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truncation according to Anderson and Ogren (1998). Due to the high particle concentration, changing the PSAP filter every time the transmission dropped below 70% (as recommended by Bond et al., 1999) was often not possible. Instead, we opted for a filter change whenever the transmission was near 50%; absorption measurements

when the corresponding transmission levels dropped below 50% were later discarded. We corrected our raw absorption coefficients from the PSAP to 550 nm for loading, apparent absorption, and multiple scattering artifacts according to Bond et al. (1999).

By alternating between total and submicron sampling with an impactor, we were able to look at PSAP absorption and nephelometer scattering for fine and coarse (total - fine) particles separately. To generate a continuous time-series of total scattering, the ratio of total to submicron scattering determined from adjacent segments was applied to each submicron scattering segment. Complete times-series of submicron scattering and total/submicron PSAP absorption were generated analogously.

To cover a greater range of wavelengths than the PSAP, we used a Magee Scientific (Berkley, CA) AE31 aethalometer to record absorption sequentially at 370, 470, 520, 590, 660, 880, and 950 nm in 2-min cycles. The aethalometer was operating in an automated mode, under which the filter tape advances when the attenuation at 370 nm reaches 75. The flow rate of the aethalometer was kept near the lower limit required by the instrument (~1.2 LPM) due to the high concentration of particles.

Following calibration schemes from Bond et al. (1999), Weingartner et al. (2003), and Arnott et al. (2005), we corrected absorption coefficient from the aethalometer for the loading artifact of absorbing particles, apparent absorption by purely scattering particles, and multiple scattering by the filter matrix using the following equation:

$$b_{ap}(t, \lambda) = (b_{aeth}(t, \lambda) - F_{aa}(\lambda) \cdot b_{sp}(t, \lambda)) \frac{F_{load}(t, \lambda)}{F_{ms}(\lambda)} \quad (3)$$

At a given time (t) and wavelength (λ), the corrected absorption coefficient (b_{ap}) is related to raw absorption coefficient from the aethalometer (b_{aeth}) by three wavelength-dependent correction factors. F_{aa} is the apparent absorption correction factor adopted

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from Arnott et al. (2005). F_{ms} is the multiple scattering correction factor taken from Weingartner et al. (2003) for coated soot.

To develop F_{load} , the loading correction factor, we first followed the parameterization from Weingartner et al. (2003) and assumed that raw absorption coefficient from the aethalometer is linearly related to the log of attenuation (ATN). When the filter tape gets heavily loaded, ATN correspondingly increases while b_{aeth} artificially decreases until the next tape advancement, resulting in discontinuities across filter changes usually observed when the aerosols are relatively constant (e.g. at ~23:00 in Fig. 1). On time-scales of minutes, variability in ambient absorption is mostly due to concentration change (driven by atmospheric dynamics) rather than composition change (driven by chemistry) in the aerosols. We can remove much of the variability in absorption by taking into account the change in concurrent scattering from the nephelometer (b_{sp}). By assuming that the remaining variability in absorption (due to aerosol composition change) is negligible (i.e. constant ω_0) on short time-scales, we can then apply the proportionality between b_{aeth} and $\text{Ln}(ATN)$:

$$F_{load}(t, \lambda) = (R_f(\lambda) - 1) \frac{\text{Ln}(ATN(t, \lambda)) - \text{Ln}(ATN(A, \lambda))}{\text{Ln}(ATN(B, \lambda)) - \text{Ln}(ATN(A, \lambda))} \quad (4)$$

Here A and B designate the first point after an aethalometer tape advancement and the last point before that advancement, respectively. At a given wavelength, R_f is defined as R_a/R_s , where $R_a = b_{aeth}(A)/b_{aeth}(B)$ and $R_s = b_{sp}(A)/b_{sp}(B)$. Scattering from the nephelometer was fitted to the seven wavelengths of the aethalometer using the scattering Ångström exponent, which was determined from the slope of $\log(b_{sp})$ vs. $\log(\lambda)$. A loading correction without taking scattering into account (i.e. forcing $b_{aeth}(A)$ and $b_{aeth}(B)$ to be the same) is only valid when the aerosols are completely non-changing and should not be applied to ambient measurements (e.g. at ~22:00 in Fig. 1). The quality of Eq. (4) was assessed by plotting the respective $b_{ap}/b_{ap}(A)$ against $\text{Ln}(ATN)$, which resulted in a largely ATN -independent ratio of unity on a campaign average. Please refer to Yang (2007) for a much more detailed description of the aethalometer correction and also a comparison of this correction scheme to ones previously pub-

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lished. Corrected absorption from the aethalometer adjusted to 550 nm by the absorption Ångström exponent showed excellent agreement with absorption from the PSAP (within 5%).

2.2 Elemental and organic carbon measurements

5 We measured elemental carbon and organic carbon using a Sunset Laboratory Inc. (Forest Grove, OR) Semi-continuous OCEC Carbon Aerosol Analyzer. Samples were collected and analyzed hourly, except 07:00–08:00 and 19:00–20:00, when filters were changed and calibrations were run. They started on the hour and lasted 40 min, allowing 20 min for analysis. A passive diffusion denuder was placed before the OCEC
10 instrument to remove any vapor-phase organic carbon. The instrument volatilizes particulate carbon thermally and oxidizes it to carbon dioxide (CO₂), which is monitored by a non-dispersive infrared (NDIR) detector. Since a portion of OC is charred to an absorbing EC-like material instead being directly oxidized to CO₂, the instrument relies on the change in transmission of a laser through the filter to account for the pyrolysis of (originally) non-absorbing OC. The split-point between EC and OC is defined
15 as the time when the laser transmission returns to its initial level (after the charred material is burnt off). This determination of the split-point is based on the assumption that non-native and native EC have the same thermal stability and optical properties, which might not be true if any liquid-like brown carbon was present (Subramanian et al.,
20 2007). Interested readers may refer to Huebert et al. (2004) and Kline et al. (2004) for more discussions on combustion stages, blanking procedures, and the determination of the split-point.

We also obtained size-segregated concentrations of EC and OC from an MSP (Minneapolis, MN) Micro-orifice Uniform Deposit Impactor (MOUDI) during selected twelve-hour intervals. The MOUDI was set to collect particles at a flow rate of 30 L/min on nine successive impactor stages. Those nine stages consisted of aluminum substrates, while a backup quartz filter in the collected the smallest particles. The aerodynamic size-cuts were 18, 10, 6.2, 3.1, 1.8, 1.0, 0.56, 0.30, 0.17, and 0.05 μm, where the

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first is a crude approximation for the inlet size, and the last is an arbitrary number for the backup filter. Particles collected on each stage were then analyzed by a similar Sunset Lab OCEC procedure, as batch samples. Because the aluminum substrate is not transparent to the OCEC laser, the split-point between OC and EC from the laser transmission could not be obtained directly for the first nine stages; instead only total carbon (EC + OC) was outputted. The final quartz filter is transparent to the laser; from that we determined the split-point between OC and EC as above and assigned it to the preceding aluminum stages, assuming that carbonaceous particles of different sizes evolve under the same temperatures. We also checked split-points obtained directly from the Sunset lab semi-continuous OCEC analyzer at concurrent times as the MOUDI; the semi-continuous instrument generally yielded comparable splits.

2.3 Inorganic particulate measurements

Every three hours, we collected particles on a 47 mm diameter Savilex Teflon filter at a flow rate of ~20 SLPM for the first two hours and 40 min (except the 05:00 to 06:40 and 17:00 to 18:40 samples that were over an hour and 40 min). After drying and weighing for the gravimetric mass, filters were extracted with water and concentrations of sodium (Na^+), potassium (K^+), chloride (Cl^-), sulfate (SO_4^{2-}), nitrate (NO_3^-), oxalate ($(\text{COO})_2^{2-}$), calcium (Ca^{2+}), magnesium (Mg^{2+}), and ammonium (NH_4^+) were measured by ion chromatography. Analytical details can be found in Kline et al. (2004).

2.4 Particle number measurements

We measured the size-distributed particle number concentration with a TSI Aerodynamic Particle Sizer (APS) every 20 s. The APS separates aerosols according to their aerodynamic (flow) diameters between ~0.5 and ~20 μm (Ananth and Wilson, 1988). However, the sizing and detecting efficiencies are relatively poor below ~0.7 μm , while significant particle loss from the inlet to the detector limits its sampling efficiency for aerosols greater than ~10 μm in diameter. For non-spherical (and non-standard den-

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sity) particles, we first applied an empirical Stokes correction according to the APS manual, and then converted aerodynamic diameters to (equivalent) geometric diameters by dividing by the square root of an estimated effective density. We used an effective density of 2.0 g cm^{-3} , which is lower than the likely actual dust density of 2.6 g cm^{-3} . Size-distributed aerosol number concentrations were converted to volume concentrations using these derived geometric diameters; mass concentrations were then calculated from volume using the likely density.

The total particle number concentration was measured by a TSI Condensation Particle Counter 7610 (CPC) every second.

2.5 Other available measurements

Basic meteorological measurements were made at a tower $\sim 75 \text{ m}$ away from the IAP building every 10 min at five levels (2, 4, 8, 16, and 32 m above ground). Our inlets on the roof of the IAP building corresponded to a height between 16 and 32 m.

Russell Dickerson's group from the University of Maryland measured gas concentrations of carbon monoxide (CO), nitric oxide (NO), reactive odd nitrogen (NO_y), and sulfur dioxide (SO_2) on 10-min intervals.

3 Results

In general, atmospheric dynamics can account for most of the observed variability in aerosol concentrations. Correlation coefficients calculated between any pairs of gaseous and/or particulate species were almost always greater than 0.6. For species closely related in origin, such as between OC and EC, the correlation was close to 0.9. Figure 2 shows the time-series of aethalometer absorption coefficients and nephelometer scattering coefficients at 520 nm (10-min averages, top panel) and EC and OC concentrations (3-h averages, bottom panel).

To examine aerosols in Xianghe over the course of an average day, we binned our

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measurements to the 24 h of a day (Fig. 3). We saw a bimodal distribution for most extensive variables, including absorption, scattering, and concentrations of OC and EC. The single scatter albedo was the lowest and the absorption Ångström was the closest to one at times when extensive properties peaked, confirming that soot carbon was one of the principal aerosols emitted from morning and evening cooking and traffic emissions.

3.1 Identification of end-member air masses

During our three weeks of sampling in Xianghe, we encountered air dominated by fresh chimney plumes, other coal-burning emissions, and occasional biomass burning and dust aerosols, often at separate times. Based on chemical, optical, and particle size measurements, we identified periods when each type of air mass was prevalent. A series of criteria was developed to classify each of these air mass types. Most criteria were of intensive or intrinsic characteristics of the aerosols and do not depend on absolute concentrations. The concentration of carbon monoxide (CO), a relatively long-lived gas (lifetime of months, Novelli et al., 1992), was usually used to normalize the concentration of the species of interest to minimize the influence of atmosphere dynamics and dilution. To stay consistent with our impactor for the PSAP and nephelometer, we defined one micron as the cut-off diameter between fine and coarse particles for the APS. Whenever possible, we tried not to use the Ångström exponent and single scatter albedo in our air mass identification scheme, since our goal was to determine \tilde{A} and $\omega_0(\lambda)$ for each air mass type. In general for each criterion, to exclude mixed air masses, we looked for maxima or minima one standard deviation or more away from the median. The median was preferred over the mean here, because the latter tended to be influenced more by sudden and extreme variations. For each type of air mass, multiple criteria needed to be met at the same time in order for the identification to be valid; if one or more of the exclusions were violated, the air mass was no longer deemed an end-member. The classification scheme is summarized in Table 1.

The identification of dust was based on particle size and composition. Calcium is one

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of the major components of these coarse particles (Arimoto et al., 2004). High Ca:CO ratios corresponded nicely with low fine scattering fraction (FSF, or the fraction of total scattering due to submicron particles from the nephelometer), as shown in Fig. 4. Low coarse number fraction (total supermicron particle number from the APS divided by the CPC count) and ω_0 were excluded since both imply a combustion source for the aerosols.

A widely used proxy for biomass burning aerosols is soluble potassium. Yet our potassium peaks did not correspond to any observed biomass burning events, possibly because sporadic agricultural fires reached our site for usually only a few minutes at a time. Those relatively weak and short-lived signals could easily be obscured in the 3-h filter samples. Instead, we looked for high single scatter albedo at 550 nm around occasions when we observed field fires upwind, which tended to give out whitish smoke and raise ω_0 when reaching our inlets. Previous studies also suggested that aerosols from biomass burning are characterized by relatively high single scatter albedo in the mid-visible (Kaufmann et al., 1992; Penner et al., 2001). Since mineral dust could also have high ω_0 at 550 nm, periods already classified as dust were not double-counted as biomass burning.

Rapidly varying chimney plumes from nearby sources were of interest because fresh emissions from these industrial boilers (likely burning diesel oil or coal) are expected to resemble “pure” soot the most. While NO and CO are both byproducts of fossil fuel combustion, NO has a very short life time in the atmosphere (hours) and its concentration is expected to fall off more quickly away from the source of combustion than the more stable CO. Classification based on this criterion was consistent with our visual and olfactory observations of chimney plumes, especially at low wind speeds.

In addition to fresh chimney plumes, there were very often long-lasting periods with constantly high levels of coal burning pollution, which likely originated from residential heating and cooking and distant smokestacks. Under stagnant conditions, these highly absorbing pollutants accumulated to very high levels. We looked for a very high SO₂:CO₂ ratio because SO₂ is one of the principal gases emitted from coal burning

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and has a much shorter lifetime than CO. Coal-derived emissions from this project generally had a relatively low ω_0 and an abundance of large particles, though neither was used as a criterion.

To identify background air, we decided to simply look for times that were low in scattering instead of using intensive properties. A high wind speed was chosen because strong winds often dissipated local pollution. We discarded times when the coarse volume fraction (fraction of total aerosol volume due to supermicron particles from the APS) was greater than 0.94 as a way to avoid double counting with mineral dust, which had already been categorized.

3.2 Optical characteristics of end-member air masses

Figure 5 shows the distinct Ångström exponents in absorption and scattering for these identified end-member air masses. The scattering Ångström is usually inversely correlated with particle size, while a high absorption Ångström exponent implies the presence of light absorbing aerosols other than BC, such as dust and brown carbon.

The means and standard deviations (in parenthesis when appropriate) of the single scatter albedo for the entire project and also for end-member air masses are summarized in Table 2. Total absorption from the aethalometer and scattering from the nephelometer were used in the calculations in 2a, whereas absorption from the PSAP and scattering from the nephelometer were separated into fine and coarse modes in 2b. The submicron and supermicron ω_0 averaged to 0.82 and 0.87 at 550 nm, bracketing that of total aerosols. As expected, submicron ω_0 was lower than supermicron ω_0 because most of the absorption was due to fine mode carbonaceous aerosols. From uncertainties in the nephelometer scattering ($\sim 7\%$, mostly due to the truncation correction) and in the aethalometer absorption (19–26% from 370 to 950 nm, mostly due to the multiple scattering correction), we estimated the uncertainties of the single scatter albedo to be about 0.03. This uncertainty is relatively consistent and systematic at all wavelengths, so we have more confidence in the spectral shapes than their absolute magnitudes.

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Table 3 shows the absorption and scattering Ångström exponents for the entire project and for identified air masses. The wavelength-dependence in absorption was calculated over the entire aethalometer wavelength range and also over the visible spectrum only, leading to almost identical results. The wavelength-dependence in scattering was calculated over the more limited nephelometer wavelengths. Both absorption and scattering Ångström exponents averaged to ~ 1.5 . The fine absorption fraction (the fraction of total absorption due to submicron particles from the PSAP) and fine scattering fraction averaged to 0.83 and 0.76, respectively, at 550 nm (Table 4).

The observed single scatter albedo was at its highest in the entire project during dust events at 0.90 (550 nm) and it increased from the UV to the IR. During EAST-AIRE, ω_0 was much lower than the 0.97 seen during ACE-Asia (Anderson et al., 2003). One possible explanation is that the Xianghe dust had included some light absorbing carbon (from interactions in the coal boiler and atmosphere, or from re-suspension of soot deposited on the ground), whereas the free tropospheric dust observed during ACE-Asia did not have much contact with urban pollutants. Submicron and supermicron ω_0 at dust-dominated times were 0.86 and 0.94 at 550 nm; both were higher than their respective project means. If only looking at total absorption and scattering, a high absorption Ångström exponent (1.89) and a low scattering Ångström exponent (0.59) were also observed at these times. The wavelength-dependence in total absorption during dust events is not as high as what was previously reported for dust only ($\text{Å} \approx 3$), presumably due to the aforementioned BC that follows a lower wavelength-dependence. The low wavelength-dependence in scattering is characteristic of coarse particles that scatter in the geometric regime. The fine absorption and scattering fractions were 0.69 and 0.46 for dust-dominated times, respectively; both were much lower than their project averages, again indicating the presence of coarse particles.

The single scatter albedo was much lower in fresh chimney plumes. Overall, ω_0 was 0.83 at 550 nm, while submicron and supermicron ω_0 were 0.79 and 0.86, respectively. The wavelength-dependence in total absorption reached its project minimum in chimney plumes at $\text{Å} = 1.35$, implying that these aerosols were similar to fresh soot. The fine

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absorption and scattering fractions were 0.85 and 0.79, respectively; both were higher than the project means, confirming that there was a large fraction of fine absorbing particles. However, this absorption Ångström exponent is still ~30% higher than what is expected for soot only (unity), which indicates the ubiquitous presence of other light absorbing aerosols near Xianghe.

Residential combustion of coal briquettes seemed to result in emissions of large particles, presumably either from the clay binder or directly from the coal, in addition to soot. The lowest ω_0 during the entire project was found in coal-derived aerosols, with an average of 0.80 at 550 nm. Unlike in fresh chimney plumes where ω_0 decreased with wavelength, ω_0 for these coal-derived aerosols showed no obvious wavelength-dependence. Compared to chimney plumes, the higher absorption Ångström (1.47) and lower scattering Ångström (1.39) in these coal-derived aerosols indicate more coarse (absorbing) particles.

The background air, which was identified by low scattering, closely resembled the typical atmosphere in Xianghe in terms of the single scatter albedo. The low ω_0 in northern China even when the atmosphere was relatively clean underlines the importance of coal burning emissions.

4 Discussion

As a first order approximation for the EC mass absorption efficiency, we can normalize total absorption by the mass concentration of EC. The resultant “apparent” EC MAE (project mean of $11.3 \text{ m}^2/\text{g}$ at 550 nm), however, would be ~40% higher than that recommended by Bond and Bergstrom (2006) for fresh soot carbon at this wavelength ($7.5 \pm 1.2 \text{ m}^2/\text{g}$). This suggests that total light absorption cannot be explained by soot carbon alone, a likelihood reinforced by the strong wavelength-dependence ($\text{Å} \approx 1.5$). Rather, other light absorbing materials were present also and absorbed a lot more at shorter wavelengths than at longer wavelengths. Brown carbon and dust are two potentially significant light absorbers that are known to have high absorption

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Ångstrom exponents in the visible spectrum. It is worth mentioning here that if brown carbon causes an artifact in thermal-optical methods, as suggested by Subramanian et al. (2007), some of the EC values might be erroneously low, leading to high “apparent” EC MAE.

5 We also considered the effect of coating of soot by transparent condensates. Recent modeling studies (Fuller et al., 1999; Schnaiter et al., 2003; Bond et al., 2006) suggested that external mixing of BC with weakly-absorbing materials does not significantly increase absorption, but a complete encapsulation of a BC core by condensates could result in an absorption enhancement of about 30~50%. While such an increase
10 would be large enough to explain our high “apparent” EC MAE, we have no reason to attribute the high absorption Ångstrom exponent of ~ 1.5 to this encapsulation effect. Modeling and laboratory studies have not indicated any consistent increase in the wavelength-dependence of BC absorption due to coating. On the contrary, Schnaiter et al. (2005) reported a slightly decreasing absorption Ångstrom exponent with increasing
15 internal mixing of black carbon with secondary organic aerosols.

Soot, being in the Rayleigh regime, is expected to have a greater wavelength-dependence in scattering than in absorption. Thus, its single scatter albedo should decrease with increasing wavelength. Both dust and brown carbon, however, can have greater wavelength-dependences in absorption than in scattering. The single scatter
20 albedo for these two types of aerosols should therefore increase with wavelength. During EAST-AIRE, the project average single scatter albedo was low in IR and UV and peaked in the mid-visible, again suggesting a mixture of BC with other absorbing aerosols.

4.1 BC Absorption

25 Accepting that BC follows the inverse-wavelength relationship from 370 to 950 nm, we can approximate BC absorption at any wavelength within this spectrum if we know its absorption at one wavelength. We can further assume that BC is the only significant light absorber at 950 nm, since both brown carbon and dust absorb weakly in the

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near-IR. Then we can use the absorption Ångström exponent of one to extrapolate BC absorption from 950 nm to shorter wavelengths. A similar argument was used in Clarke et al. (2007), but based on negligible non-soot absorption at 670 nm.

We define the difference between total absorption and extrapolated BC absorption as residual absorption. The residual absorption was zero at 950 nm by definition and increased to over 30% of total absorption at 370 nm. In the mid-visible, the contribution to total absorption by aerosols other than BC was ~15%. The residual absorption had an absorption Ångström exponent of ~3.5 in the visible spectrum, which is close to those of brown carbon and dust. The very high correlation coefficient between residual absorption and OC concentration (0.95) suggested that most of the residual absorption was due to brown carbon.

We can now normalize the extrapolated BC absorption to the elemental carbon concentration to obtain the EC mass absorption efficiency. The resultant “true” EC MAE (project mean of 9.5 m²/g at 550 nm) represents a ~20% reduction from the “apparent” EC MAE, and is much closer to the recommendation by Bond and Bergstrom (2006) and well within the range of enhancements likely to be caused by encapsulation.

4.2 Dust absorption

While the absorption of mineral dust depends on its content of ferric oxides (Sokolik and Toon, 1999; Alfaro et al., 2004), in urban environments, volatile materials such as OC, sulfates, and nitrates from combustion can be adsorbed onto dust surfaces (Falkovich et al., 2004; Zhang and Carmichael, 1999), possibly modifying the dust optical properties. From size-segregated mass concentrations sampled by the MOUDI, we see that unlike pollution times when most mass was concentrated in the fine mode, dust events were characterized by a high fraction of mass, including that of carbonaceous particles, in the coarse mode (Fig. 6). Dust particles sampled in polluted places like Xianghe, therefore, should not be considered pure mineral aerosols. During EAST-AIRE, dust events usually coincided with high winds which re-suspended local coarse particles; it is unclear whether long-distance transport of desert aerosols was impor-

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tant. As mentioned before, some of these coarse particles might be generated during the combustion of coal (briquettes). This scenario is more plausible in bringing light absorbing carbon into the coarse mode than natural coagulation.

Since we could not directly measure the mass concentration and light absorption of dust, we estimated these two quantities from the size-distributed number concentrations from the APS using Mie scattering theory. Knowing that pollution aerosols tend to dominate in the fine mode, we only used the supermicron portion of the size-distributed number concentrations for this Mie calculation. Even though the dust number distribution might have a tail in the fine mode, (which we could not constrain due to a lack of size-distributed chemical concentration measurement and particle number count below $\sim 0.7 \mu\text{m}$) mass, scattering, and absorption due to dust are all dominated by coarse particles, as shown by Clarke et al. (2004) in their measurements during ACE-Asia. This Mie calculation also required a reasonable refractive index for dust aerosols. However, the imaginary component of the refractive index of dust varies over an order of magnitude among literature values, from 0.0006i to 0.0015i to 0.008i in the mid-visible (Clarke et al., 2004; Haywood et al., 2003; World Meteorological Organization, 1986).

To determine the refractive index of dust appropriate for Xianghe, we first focused on identified end-member dust events during EAST-AIRE and assumed that absorption due to brown carbon was negligible at those times (i.e. residual absorption, or the difference between total absorption and BC absorption, was solely due to dust). This assumption was supported by the fact that correlation between coarse volume and residual absorption at 370 nm was ~ 0.9 during dust events, much higher than ~ 0.5 for the project average. At each aethalometer wavelength, we varied the imaginary component of the refractive index from 0i to 0.01i, while keeping the real component constant at 1.53 – a value commonly used to model dust (Sokolik and Toon, 1999; Haywood et al., 2003; Clarke et al., 2004). Absorption predicted by Mie scattering theory was then summed over all size-bins to generate total absorption. The fitting was deemed optimal when the square of the difference between residual absorption and Mie-predicted absorption was the smallest on average at those end-member dusty

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periods. For 370, 470, 520, 590, 660, 880, and 950 nm, our least-square-fit method yielded an imaginary component of 0.0056i, 0.0033i, 0.0026i, 0.0019i, 0.001i, 0.0001i, and 0i, respectively. By interpolation, a refractive index of 1.53–0.0023i at 550 nm was obtained, which is in the range of previously published results. Now for the entire EAST-AIRE project, using Mie theory, we applied this set of empirically-derived refractive indices to all supermicron particle number concentrations to obtain a complete time-series of dust absorption. This calculation assumes that all coarse particles during EAST-AIRE had the same optical properties, whether during dust events or not, which might not be valid if some of the large aerosols were combustion-origin. Based on this estimation, absorption due to dust was usually quite minor in Xianghe except during dust events.

Considering again only the supermicron particles from the APS, we calculated the mass concentration of dust from the volume concentration using a constant density of 2.6 g/cm³. We could also estimate the mass concentration of dust from soluble calcium, as previous experiments suggested that water-soluble calcium makes up ~5–8% of total dust mass (Arimoto et al., 2004); this latter approach yielded similar results. The mean dust MAE was determined to be 0.03 m²/g at 550 nm, over two orders of magnitude less than that of black carbon. Published results of the absorption Ångström exponent of dust are over the visible spectrum only, likely because a power law fit might not be suitable for dust absorption from 370 to 950 nm due to its particular spectral shape, as seen in modeled spectra from Sokolik and Toon (1999). We calculated our wavelength-dependence in dust absorption from 470 to 660 nm (Table 5), resulting in an average of 3.8. A large range in the wavelength-dependence was observed, which could be due to a number of modification processes that these dust aerosols might have undergone.

4.3 Brown carbon absorption

We derived our absorption due to brown carbon by subtracting BC and dust absorption from total absorption at all wavelengths, assuming that nothing besides black carbon,

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brown carbon, and dust had any significant absorption between 370 and 950 nm during EAST-AIRE. For the mass concentration of brown carbon, we decided to simply use the total particulate organic carbon concentration measured by the Sunset lab OCEC Semi-continuous analyzer. Knowing that a large fraction of organic carbon is non-absorbing, dividing brown carbon absorption by an upper-limit concentration will give us a lower-limit mass absorption efficiency. The real mass absorption efficiency for brown carbon will surely be higher than the values presented here. The advantage of normalizing brown carbon absorption to organic carbon, however, is that the mass concentration of organic carbon can be measured by the same technique in different labs (Schauer et al., 2003) and is thus more reproducible than that of brown carbon.

The mean brown carbon MAE was determined to be $0.5 \text{ m}^2/\text{g}$ at 550 nm, less than one tenth of the EC MAE at the same wavelength. The average absorption Ångstrom exponent of brown carbon during EAST-AIRE was 3.5 over the visible spectrum, which is higher than what were commonly reported for biomass burning aerosols ($\text{\AA} \approx 2$; Clarke et al., 2006; Kirchstetter and Novakov, 2004; Kline, 2004). However, when contribution from BC was removed from total absorption, those authors found similarly high absorption Ångstrom exponents.

Table 5 summarizes the project mean mass absorption efficiencies (m^2/g) and absorption Ångstrom exponents of black carbon, brown carbon, and dust. Via propagation of errors, the uncertainty in the EC concentration from the OCEC analyzer was estimated to be $\sim 17\%$ (Kline, 2004). Combined with the propagated uncertainties in the aethalometer absorption of $19\sim 26\%$, the EC mass absorption efficiency should have uncertainties of $26\sim 31\%$ (from 370 to 950 nm). Figure 7 shows the average contribution to total light absorption by black carbon, brown carbon, and dust from 370 to 950 nm. Black carbon was predictably the most important light-absorbing aerosol; though its importance to total absorption decreased towards shorter wavelengths, where brown carbon and dust became more significant. Brown carbon on average could account for $\sim 30\%$ of total absorption at 370 nm. Even at mid-visible wavelengths, it still contributed more than 10%, underlining the optical importance of

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absorbing organic carbon in environments dominated by coal burning. The absorption due to dust was generally small (less than 5% at 370 nm), though its contribution to total absorption can be similar to that from BC during dust events.

5 Conclusions

5 A variety of in situ optical and chemical measurements were made on aerosols near Xianghe, China during the EAST-AIRE campaign. Based on previously published methods, a correction scheme incorporating concurrent scattering was developed to address the loading artifact of the aethalometer. The single scatter albedo from 370 to 950 nm was directly calculated from aethalometer absorption and nephelometer scattering. In general, aerosols in Xianghe were highly absorbing, as demonstrated by the low ω_0 at all wavelengths (project mean and standard deviation of 0.84 ± 0.05 at 550 nm). We also quantified absorption and scattering in the coarse and fine modes separately at 550 nm; the submicron ω_0 was found to be lower than the supermicron ω_0 . On average, both absorption and scattering Ångström exponents were close to 1.5, suggesting the presence of large, absorbing particles.

We identified five end-member air mass types that might be characteristic of Northern China, which are dust, biomass burning aerosols, fresh (industrial) chimney plumes, other (residential and commercial) coal-derived aerosols, and the background air. Optical properties varied significantly among these air masses as a result of their different chemical compositions. In the mid-visible, dust and biomass burning aerosols had the highest single scatter albedo among identified air masses. Dust, as characterized by high calcium and coarse scattering, had the lowest scattering Ångström exponent due to the large particle size and a high absorption Ångström exponent due to the presence of ferric oxides. During dust events, a higher percentage of the carbonaceous mass was found in the coarse mode than during pollution-dominated times, suggesting a possible surface adsorption mechanism of light absorbing carbon onto dust particles. Biomass burning and coal burning aerosols both showed high absorption Ångström

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exponents, likely because brown carbon was also produced. Fresh chimney plumes contained aerosols most resembling diesel soot (absorption \AA closest to one).

We found that normalizing total absorption to the mass concentration of elemental carbon yielded an unrealistically high EC mass absorption efficiency, indicating that light absorbers other than soot were present. We were able to attribute total light absorption to black carbon, brown carbon, and dust from 370 to 950 nm, resulting in mass absorption efficiencies of 9.5, 0.5, and $0.03 \text{ m}^2/\text{g}$ at 550 nm, respectively. The black carbon MAE might have been enhanced due to the coating effect. The brown carbon MAE should be viewed as a lower-limit since we normalized its absorption to the concentration of OC. The dust MAE represented those of local coarse particles that had interacted with pollution rather than pure desert aerosols. The contribution to total absorption from BC decreased towards the UV, where absorption due to brown carbon and dust became increasingly significant as a result of their greater wavelength-dependences. These mass absorption efficiencies from the UV to the IR hopefully will prove useful for future modeling purposes and the apportionment of gross optical properties.

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Table 1. Identification criteria for end-member air masses.

Air Mass Type	Identification Criteria	Exclusions
Dust	Coarse volume fraction ^a 1σ above median FSF ^c at 550 nm 1σ below median Ca:CO ratio 1σ above median	Coarse number fraction ^b below 0.01 ω_0 at 550 nm below 0.8
Biomass burning aerosols	ω at 550 nm above 0.87 within 1 h of observed upwind agricultural fires	Times already identified as dust
Fresh chimney plumes	NO:CO ratio 1σ above median Wind speed at 32 m 1σ below median	
Other coal pollution	SO ₂ :CO ratio 2σ above median	Times already identified as fresh chimney plumes
Background air	b_{sp} at 550 nm less than 50 Mm ⁻¹ Wind speed at 32 m above 8 m/s	Coarse volume fraction above 0.94

^a The fraction of total particle volume due to supermicron particles, as inferred from the APS

^b The fraction of total number due to supermicron particles, as inferred from the APS and CPC

^c Fine Scattering Fraction: the fraction of total scattering due to submicron particles

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Table 2a. Single scatter albedo calculated from aethalometer absorption and nephelometer scattering.

	Single Scatter Albedo, Mean (Std)						
	370 nm	470 nm	520 nm	590 nm	660 nm	880 nm	950 nm
Project mean	0.82 (0.06)	0.83 (0.05)	0.83 (0.05)	0.83 (0.05)	0.83 (0.06)	0.83 (0.06)	0.82 (0.06)
Dust	0.84 (0.03)	0.87 (0.03)	0.89 (0.03)	0.90 (0.03)	0.91 (0.04)	0.93 (0.05)	0.93 (0.06)
Biomass burning	0.88 (0.01)	0.88 (0.01)	0.89 (0.01)	0.89 (0.01)	0.88 (0.01)	0.88 (0.02)	0.88 (0.02)
Fresh plume	0.83 (0.04)	0.83 (0.04)	0.83 (0.04)	0.82 (0.04)	0.82 (0.04)	0.81 (0.04)	0.81 (0.04)
Coal pollution	0.78 (0.05)	0.80 (0.05)	0.80 (0.04)	0.80 (0.05)	0.80 (0.05)	0.80 (0.05)	0.80 (0.05)
Background	0.83 (0.05)	0.84 (0.05)	0.84 (0.05)	0.84 (0.05)	0.83 (0.05)	0.83 (0.06)	0.82 (0.06)

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Table 2b. Single scatter albedo for total, submicron, and supermicron aerosols.

Single Scatter Albedo at 550 nm ^a , Mean (Std)			
	<i>Total</i>	<i>Submicron</i>	<i>Supermicron</i>
Project mean	0.84 (0.05)	0.82 (0.05)	0.87 (0.08)
Dust	0.90 (0.04)	0.86 (0.05)	0.94 (0.03)
Biomass burning	0.89 (0.01)	0.87 (0.01)	0.91 (0.02)
Fresh plume	0.83 (0.04)	0.79 (0.03)	0.86 (0.05)
Coal pollution	0.80 (0.05)	0.79 (0.04)	0.87 (0.06)
Background air	0.84 (0.05)	0.84 (0.04)	0.88 (0.08)

^a Absorption from the PSAP

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Table 3. Ångström exponents of absorption (over the entire aethalometer wavelengths and also the visible spectrum only) and scattering (over the nephelometer wavelengths).

	Ångström Exponents, Mean (Std)		
	<i>Absorption</i>	<i>Scattering</i>	
	370~950 nm	470~660 nm	450~700 nm
Project mean	1.46 (0.27)	1.47 (0.38)	1.47 (0.25)
Dust	1.82 (0.90)	1.89 (1.63)	0.59 (0.41)
Biomass burning	1.49 (0.08)	1.53 (0.10)	1.52 (0.18)
Fresh plume	1.35 (0.10)	1.35 (0.12)	1.49 (0.11)
Coal pollution	1.46 (0.15)	1.47 (0.17)	1.39 (0.20)
Background	1.50 (0.22)	1.55 (0.26)	1.58 (0.18)

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Table 4. Fine absorption fraction and fine scattering fraction.

	Fine Absorption Fraction ^a		Fine Scattering Fraction ^b	
	550 nm	450 nm	550 nm	700 nm
Project mean	0.83 (0.10)	0.80 (0.09)	0.76 (0.10)	0.69 (0.11)
Dust	0.69 (0.18)	0.53 (0.18)	0.46 (0.17)	0.36 (0.16)
Biomass burning	0.80 (0.06)	0.79 (0.05)	0.73 (0.05)	0.65 (0.06)
Fresh plume	0.85 (0.12)	0.82 (0.10)	0.79 (0.10)	0.73 (0.11)
Coal pollution	0.85 (0.08)	0.82 (0.08)	0.77 (0.08)	0.70 (0.08)

^a the fraction of absorption due to fine mode particles, Mean (Std)

^b the fraction of scattering due to fine mode particles, Mean (Std)

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Table 5. Mass absorption efficiencies and absorption Ångstrom exponents of black carbon, dust, and brown carbon.

	Mass Absorption Efficiency – Project Mean [m ² /g]						Absorption Å 470~660 nm
	370 nm	470 nm	520 nm	590 nm	660 nm	880 nm	
Total Absorption ^a	21.4	14.6	12.3	10.2	8.9	5.9	1.47
BC ^a	14.1	11.1	10.0	8.8	7.9	5.9	1.00
Dust ^b	0.087	0.050	0.037	0.027	0.013	0.001	3.80
Brown Carbon ^c	>2.2	>1.0	>0.63	>0.37	>0.32	>0.02	3.50

^a Absorption normalized by [EC]

^b Absorption normalized by estimated dust concentration

^c Absorption normalized by [OC]

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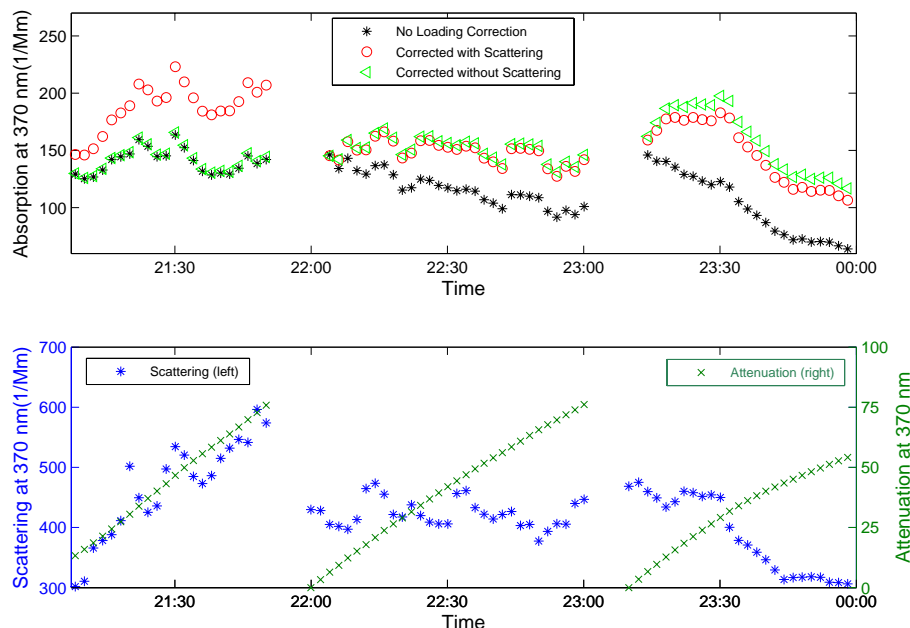


Fig. 1. Aethalometer loading correction: absorption coefficients not corrected, corrected with scattering, and corrected without scattering (top panel); scattering coefficient from the nephelometer and attenuation from the aethalometer (bottom panel). The aethalometer filter tape advances when attenuation at 370 nm exceeds 75. During the filter advancement at ~22:00, the aerosol concentration dropped rapidly, as indicated by concurrent scattering; this is reflected in the aethalometer absorption corrected with scattering, but not in the absorption corrected without scattering, where absorption before and after the filter change were forced to be equal. The aerosol concentration was relatively steady during the filter advancement just after 23:00, so corrections with and without scattering led to similar results.

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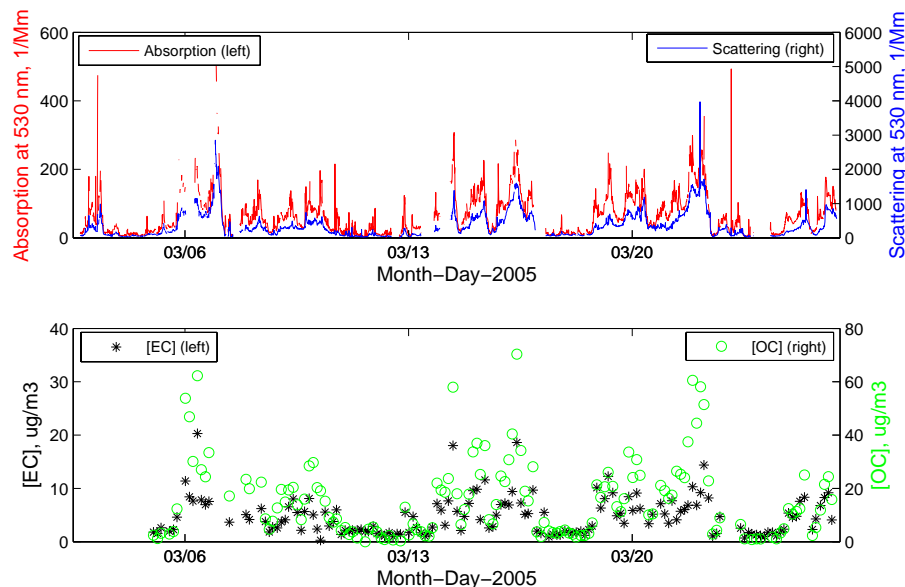


Fig. 2. Ten-minute averages of ambient absorption coefficients (1/Mm) from the aethalometer and scattering coefficients (1/Mm) from the nephelometer at 520 nm (top panel); filter-interval (~ 3 h) averages of EC and OC concentrations ($\mu\text{g}/\text{m}^3$) from the Sunset Lab semi-continuous OCEC analyzer (bottom panel). The correlation coefficient between EC and OC was about 0.9.

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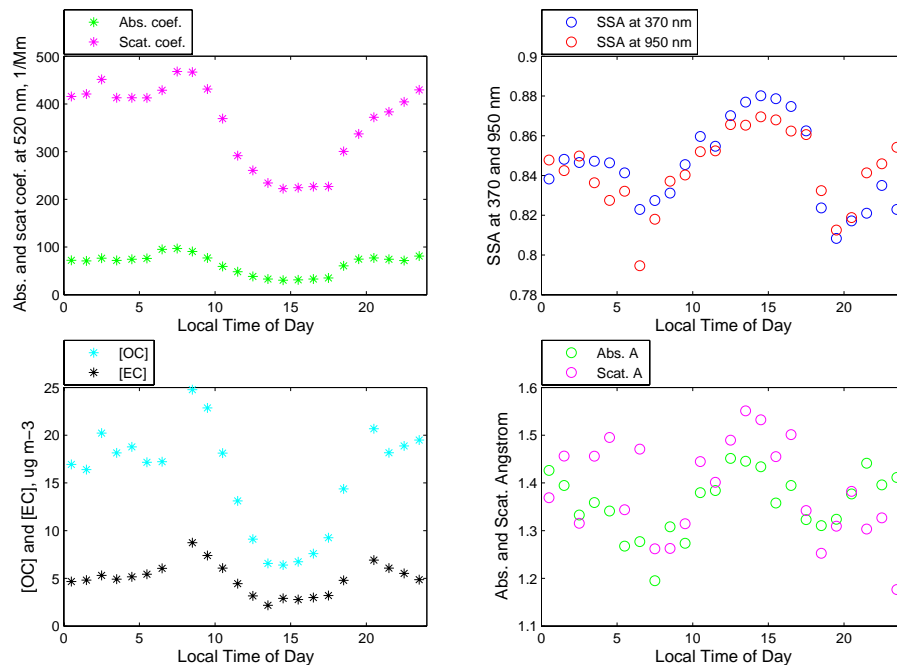


Fig. 3. Absorption and scattering at 520 nm (top left panel), single scatter albedo at 370 and 950 nm (top right panel), [OC] and [EC] (bottom left panel), and absorption and scattering Ångström exponents (bottom right panel) binned to the 24 h of a day. Extensive properties (left column) increased in morning and evening traffic, cooking hours, whereas intensive properties (right column) decreased, suggesting that a large fraction of those emissions was soot carbon.

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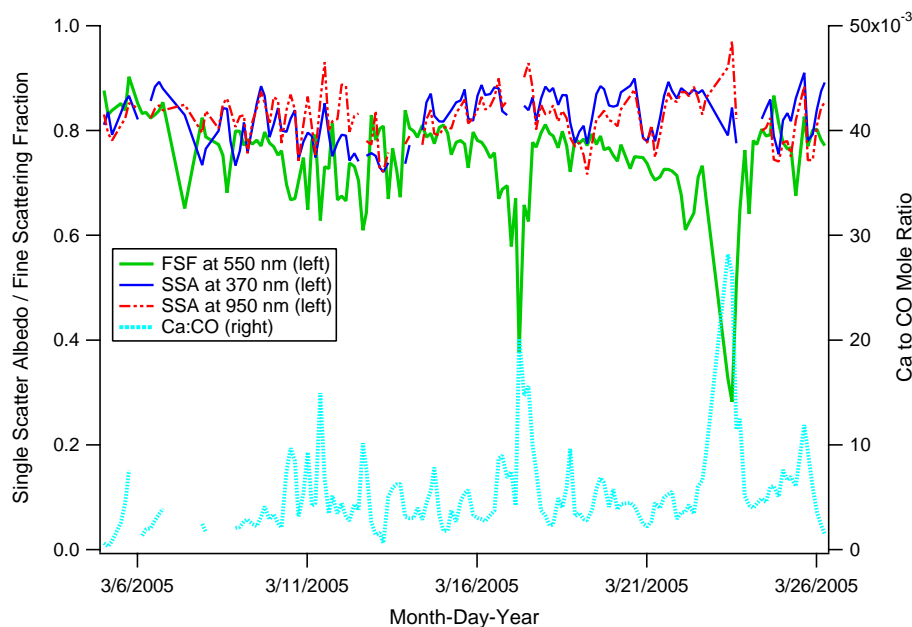


Fig. 4. Identification of dust by size (low fine scattering fraction, FSF) and soluble calcium (high Ca:CO ratio). Notice that the single scatter albedo decreased at 370 nm and increased at 950 nm during dust events (e.g., 23 March) as a result of the high absorption Ångström exponent of dust.

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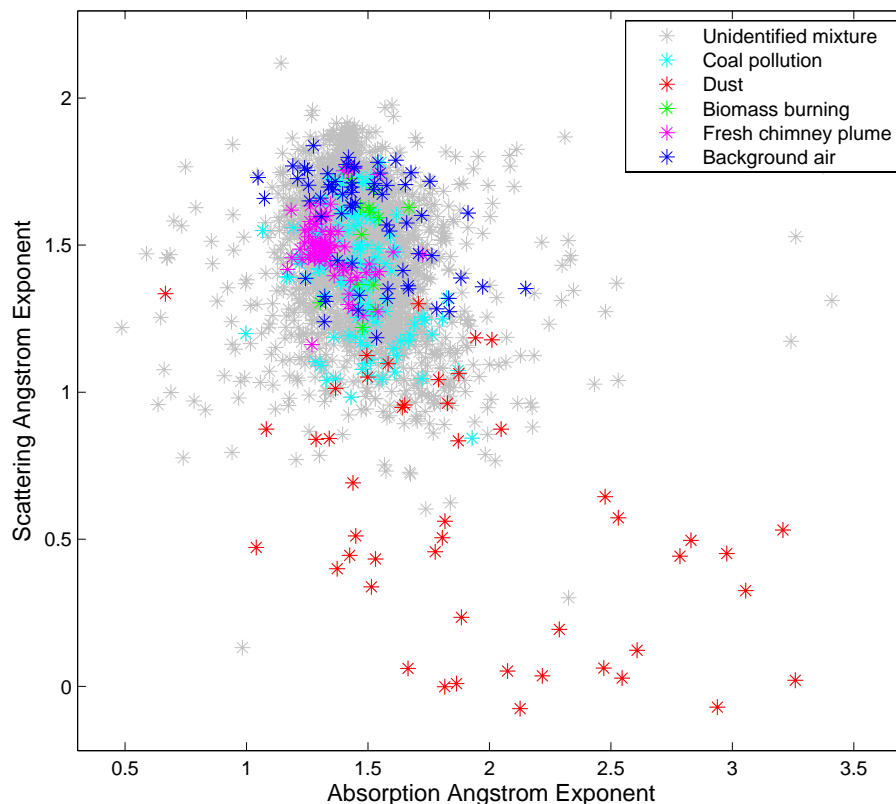


Fig. 5. Grouping of air masses by absorption and scattering Ångström exponents. Dust generally had a low scattering Å (close to zero) due to the large particle size, while its absorption Å was high but variable. Fresh chimney plumes were characterized by a near-unity absorption Å, which is expected for soot carbon. In comparison, coal pollution aerosols were much more variable in optical characteristics, likely due to formation of brown carbon and clay in addition to soot carbon.

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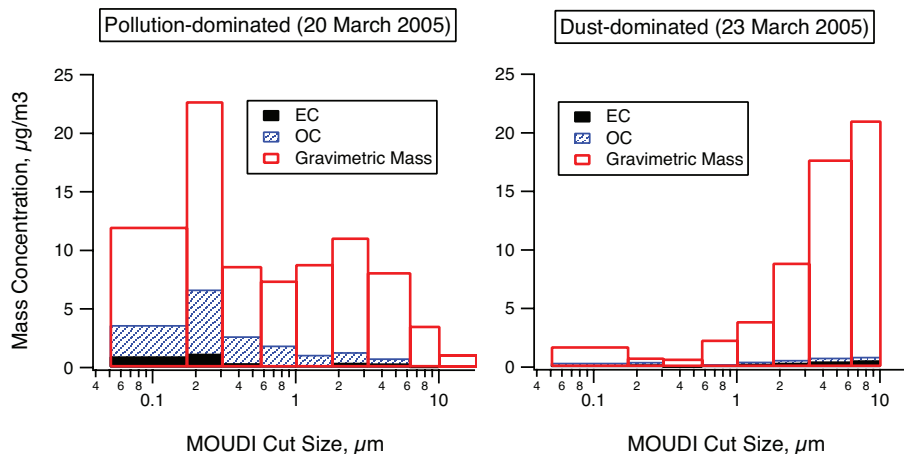


Fig. 6. Size-segregated mass concentrations from the MOUDI. Most carbonaceous mass (OC and EC) was in the fine mode during pollution-dominated times (left panel). However, during dust-dominated times (right panel), there was a lot more carbon (~50%) present in the coarse mode, possibly adsorbed onto surfaces of dust particles.

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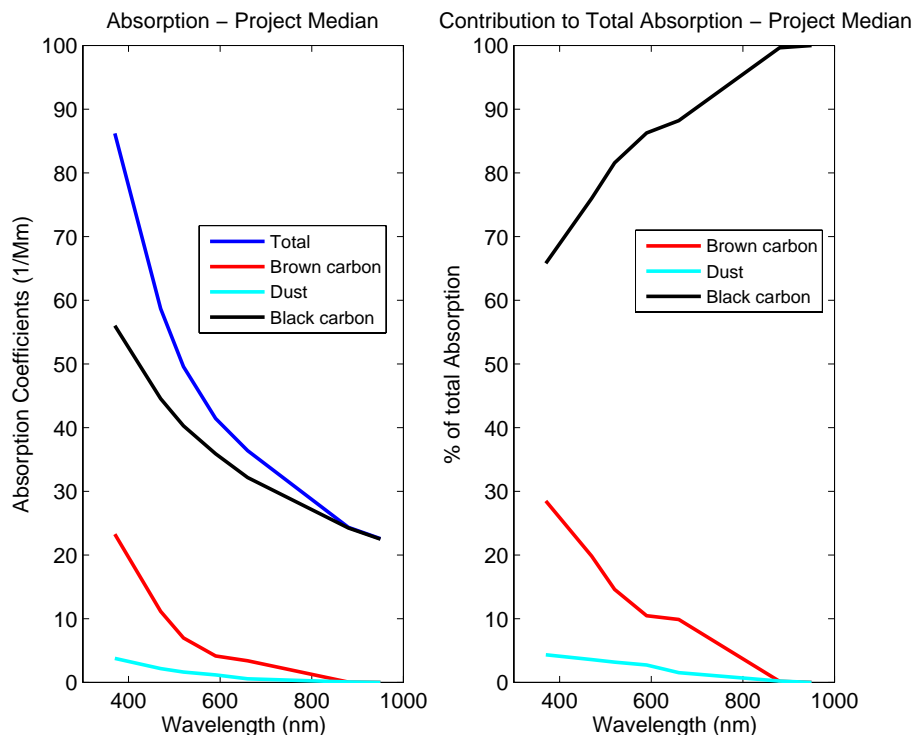


Fig. 7. Apportioning of total light absorption to black carbon, brown carbon, and dust: project medians of their absolute absorption (left panel) and relative contributions (right panel). BC's contribution to total absorption decreased towards the UV, where brown carbon and dust became more important. For this coal-dominated environment, brown carbon could account for ~30% of total light absorption in the UV, and over 10% at mid-visible wavelengths.

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